nuclear ribosomal DNA: a valuable source of evidence on angiosperm phylogeny. *Ann. Mo. Bot. Gard.,* 1995,82,247-277.

- 15. Williams, J. G. K., Kubelik, A. R., Livak, K. J., Rafalshi, J. A. and Tingey, S. V., DNA polymorphisms amplified by arbitrary primers are useful as genetic markers. *Nucleic Acids Res,* 1990, 18,6531- 6535.
- 16. Welsh, J. and McClelland, M., Fingerprinting genomes using PCR with arbitrary primers. *Nucleic Acids Res.*, 1990, 18, 7213-7218.
- 17. Heath, D. D., Iwana, G. K. and Delvin, R. H., PCR primed with VNTR core sequences yield species-specific patterns and hypervariable probes. *Nucleic Acids Res.,* 1993,21,5782-5785.
- 18. Guo, Z. H., Chen, Y. Y. and Li, D. Z., Phylogenetic studies on the *Thamnocalamus* group and its allies (Gramineae: Bambusoideae) based on ITS sequence data. *Mol. Phylogenet. Evol.,* 2002, 22, 20- 30.
- 19. Ronsted, N., Chase, M. W., Albach, D. C. and Bello, M. A., Phylogenetic relationships within *Plantago* (Plantaginaceae): evidence from nuclear ribosomal ITS and plastid trnL-F sequence data. *Bot. 1. Linn. Soc.,* 2002, 139, 323-338.
- 20. Bhattacharya, E. and Ranade, S. A., RAPD and DAMD profile differences amongst mulberry varieties. *BMC Plant BioI.,* 2001, 1, 3; http://www.biomedcentral.com/content/pdf/1471-2229-1-3.pdf
- 21. Ranade, S. A. and Goswami, M., Inter-species affinities in the genus *Prosopis* based on RAPD profiles. In Proceedings of the Symposium on Advances in Legume Research in India (ed. Rao, R. R.), Bishen Singh Mahinder Pal Singh, Dehradun, 2002, pp. 465- 475.
- 22. Verma A., Kumar, N. and Ranade, S. A., Genetic diversity amongst landraces of a dioecious vegetatively propagated plant, betelvine *(Piper betle* L.). *1. Biosci.,* 2004, 29,319-328.
- 23. Chase, M. W., Morton, C. M. and Kallunki, J. A., Phylogenetic relationships of Rutaceae: a cladistic analysis of the subfamilies using evidence from rbcL and atpB sequence variation. *Am. 1. Bot.,* 1999,86,1191-1199.
- 24. Samuel, R., Ehrendrorfer, F., Chase, M. W. and Greger, H., Phylogenetic analyses of Aurontioideae (Rutaceae) based on noncoding plastid DNA sequence and phytochemical features. *Plant BioI.,* 2001, 2001, 77-87.
- 25. Morton, C. M., Grant, M. and Blackmore, S., Phylogenetic relationships of the Aurantioideae inferred from chloroplast DNA sequence data. *Am. 1. Bot.,* 2003, 90, 1463-1469.
- 26. Navarro, F. B., Suarez-Santiago, V. N. and Blanca, G., A new species of *Haplophyllum* A. Juss. (Rutaceae) from the Iberian peninsula: evidence from morphological, karyological and molecular analyses. *Ann. Bot.,* 2004, 94, 571-582.
- 27. Rossetto, M., A simple molecular approach for identifying a rare *Acronychia* (Rutaceae) provides new insights on its multiple hybrid origin. *BioI. Conserv.,* 2005, 121, 35-43.
- 28. Scott, K. D., McIntyre, F. and Playford, J., Molecular analysis suggests a need for a significant reassessment of Rutaceae subfamilies and a minor reassessment of species relationships within *Flindersia. Plant Syst. Evol.,* 2000, 223, 15-27.
- 29. Pavlicek, A., Hrda, S. and Flegr, J., FreeTree Freeware program for construction of phylogenetic trees on the basis of distance data and bootstrapping/jackknife analysis of the tree robustness. Application in the RAPD analysis of the genus *Frenkelia. Folia BioI. (Praha.),* 1999, 45, 97-99; http://www.natur.cuni.cz/~flegr/free tree.htm
- 30. Page, R. D. M., TreeView (Win32), ver. 1.6.5, 2001; http:// taxonomy.zoology.gla.ac.uk/rod/rod.html
- 31. Kollipara, K. P., Singh, R. J. and Hymowitz, T., Phylogenetic and genomic relationships in the genus *Glycine* Willd. Based on sequences from the ITS region of nuclear rDNA. *Genome,* 1997,40, 57-68.
- 32. Altschul, S. F., Gish, W., Miller, W., Myers, E. W. and Lipman, D. J., Basic local alignment search tool. *1. Mol. BioI.,* 1990, 215, 403-410; http://www.ncbi.nlm.nih.gov/BLAST/
- 33. Morgenstern, B., DIALIGN 2: improvement of the segment-tosegment approach to multiple sequence alignment. *Bioinformatics,* 1999, 15, 211-218: http://bibiserv.techfak.uni-bielefeld.de/dialign/welcome.html
- 34. Shannon, C. E. and Weaver, W., *The Mathematical Theory of Communication,* Univ. of Illinois Press, Urbana, 1949.
- 35. Nei, M., Genetic distance between populations. *Am. Nat., 1972,* 106, 283-292.
- 36. Yeh, F. c., Yang, R.-C. and Boyle, T., POPGENE A Microsoft Windows-based freeware for population genetic analysis, ver. 1.32 (32 bit), 1997: http://www.ualberta.ca/-fcyeh/
- 37. Jobes, D. V. and Thein, L. B., A conserved motif in the 5.8S ribosomal RNA (rRNA) gene is a useful diagonostic marker for plant internal transcribed spacer (ITS) sequences. *Plant Mol. BioI. Rep.,* 1997,15,326-334.

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ATP levels and adenylate energy charge in soils of mangroves in the Andamans

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Adenosine 5' -triphosphate (ATP) is considered to be a useful indicator of life in soil and the adenylate energy charge (AEC) indicates the energetic status of soil microorganisms. A TP concentration and AEC levels have been extensively studied in a diverse group of soils. However, little knowledge is available on the levels of ATP and AEC in soils of mangroves. We report here the levels of adenylates ATP, adenosine di-phosphate (ADP) and adenosine monophosphate (AMP)) and AEC in soils of undisturbed mangroves of South-, Middle-, North- and Little-Andamans. Relevant soil physico-chemical and microbial parameters and their relationship to ATP and AEC were also examined. A veraged across various mangrove sites, total N level was 1.44 ± 0.13 g kg⁻¹, organic C 15.6 ± 1.5 g kg⁻¹, microbial biomass C 410 \pm 35 μ g kg⁻¹, microbial biomass N 34 \pm 2 μ g kg⁻¹ and qCO₂ 41.1 \pm 4.4 mg CO₂ (g bio- $\frac{1}{2}$ $\frac{1}{2}$ $\frac{1}{2}$ $\frac{1}{2}$ $\frac{1}{2}$ and $\frac{1}{2}$ $\frac{1}{2}$ $\frac{1}{2}$ + \frac

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from 2.32 to 3.22 nmol g^{-1} (mean 2.87 \pm 0.29), AMP from 0.21 to 0.29 nmol g^{-1} (mean 0.25 \pm 0.03) and ADP from 0.41 to 0.48 nmol g^{-1} (mean 0.44 \pm 0.03). Across sites, the average microbial biomass C/organic C ratio was 2.6 ± 0.2% and microbial biomass *C/N* ratio at the mangrove sites was wider and ranged from 11.2 to 14.5 with a mean of 12.0 ± 0.9 . The ATP/microbial biomass C ratio ranged from 6.0 to 8.2 μ mol g⁻¹ with a mean of 7.0 \pm 0.6 µmol g⁻¹, markedly lower than the worldwide average of $10-12 \mu$ mol g^{-1} reported in a wide range of soils. Lower ATP/microbial biomass C ratio in our mangrove soils is most likely due to a changed microbial community structure indicating a decomposition pathway dominated by fungi and microorganisms with large microbial biomass *C/N* ratio. The AEC levels were consistently >8.0 (mean 0.87) at all the sites, suggesting that the majority of microorganisms in these mangrove soils are probably dormant.

Keywords: Adenylates, ATP, ATP/microbial biomass C ratio, adenylate energy charge, mangrove forests.

ADENOSINE 5' -triphosphate (ATP) occurs in all living cells¹, but exocellular ATP has a half-life of less than 1 h. The ATP content is, therefore, considered a useful indicator of life in soil². Besides, there is substantial evidence to suggest that the soil microbial biomass maintains an ATP concentration typical of microorganisms undergoing exponential growth *in vitro*^{3,4}. However, the soil microbial populations are supposed to be predominant in a dormant state with low metabolic activity and low turnover rates⁵. It was proposed that the energetic status of soil microorganisms can be evaluated by determining the adenylate energy charge $(AEC)^6$. In cultures of microorganisms *in vitro*, AEC values > 0.8 indicate actively growing cells, values from 0.5 to 0.7 represent dormant cells that are incapable of biosynthesis, and values $\lt 0.4$ occur only in dead or dying cells⁷ . Pioneering work on adenylates (ATP, adenosine di-(ADP) and monophosphates (AMP)) and AEC in soils was done by Jenkinson and co-workers^{8,9}, as well as Brookes and co-workers^{6,10}. Subsequently, Contin *et al.*² combined appropriate published data on ATP available up to 1996 and some of their own results in addition to Jenkinson's data⁹ to reexamine the literature on ATP and microbial biomass relationships in a wide group of soils from the northern hemisphere, southern hemisphere and Japanese paddy-dryland crop rotation. Notably, this included soils under different management regimes encompassing arable, grassland and woodland soils. More recently, the literature base of ATP and biomass relationship was made larger and diverse by Joergensen and co-workers^{4,11-13} through their excellent work on a wide group of soils of temperate and tropical ecosystems. Various other published data on adenylates in soils also exist $14-17$.

However, information on adenylates especially ATP and AEC in soils under the mangroves is limited. Mangroves are one of the most unique and endangered ecosystems of the biosphere covering 60-70% of the tropical coasts, especially in India, Thailand, the Philippines, Malaysia, Indonesia, Bangladesh and Papua New Guinea. In India, mangroves occur mostly in the west coast (Kerala, Karnataka, Goa, Maharashtra, Gujarat, coral atolls of Lakshadweep islands), and east coast (Tamil Nadu, Andhra Pradesh, Orissa, West Bengal, and the Andaman and Nicobar Islands). Among these, mangroves of the Andamans are considered to be the most luxuriant¹⁸, covering about 77,769 ha 18 .

For the study, four undisturbed mangroves sites were selected from each district of the Andamans (10°30'- 13°42'N lat. and 92°l4'-94°l6'E long.) and 20 random cores (0-15 cm, 7 cm \varnothing) were taken from each site. The soils were then sieved $(< 2$ mm), and analysed for their moisture content. Sub-samples for the determination of organic carbon and total N were sieved to pass through a 0.5 mm mesh. Soil pH was determined in a 1: 2.5 soil: water suspension, organic C by the Walkley Black method¹⁹, total N by the Kjeldahl method²⁰, clay content by the pipette method²¹ and cation exchange capacity (CEC) by the method of Gillman²². The microbial biomass C and N were estimated by fumigation-extraction²³ using a factor of 0.45 (ref. 24) and 0.54 (ref. 25) respectively. The adenylates (ATP, AMP and ADP) were estimated by the procedure of Dyckmans and Raubuch²⁶. Dimethylsulphoxide (DMSO), $Na₃PO₄$ -buffer (10 mM), EDTA (20 mM) and a nucleotide-releasing buffer (benzalkonium chloride containing 2 mM Mg-EDTA, 10 mM ammonium acetate and 20 mM THAM, pH 7.75 with acetate)¹¹ were used as extractants. The energy status of soil microorganisms was evaluated by determining AEC, which is defined as: $AEC = (ATP + 0.5 \times ADP/(ATP +$ ADP + AMP)²⁷. The metabolic quotient (qCO₂) was determined by measuring basal respiration $(CO₂)$ evolution) in moist soil samples adjusted to 55% of its water-holding capacity. Briefly, the samples were pre-incubated for 3 days at 20 $\rm ^{o}C$ in the dark followed by measuring $CO₂$ production for another 3 days by trapping $CO₂$ in 0.05 M NaOH. $CO₂$ production was then measured by titration of the excess NaOH with 0.05 M HCI. The metabolic quotient was calculated using the formula: $(\mu gCO_2-C$ evolved in 3 days g^{-1} soil)/(μ g biomass C g^{-1} soil)/3 days \times $1000 = mg CO₂-C g⁻¹$ biomass C per day¹². All values reported are means of 20 determinations expressed in an oven-dry basis (24 h at 105°C).

The results (Table 1) revealed that soil pH varied in a narrow range of 5.20-6.05, clay between 19 and 27%, CEC between 212 and 268 μ mol g^{-1} , total N between 1.31 and 1.83 g kg⁻¹ and organic C between 13.9 and 19.8 $g \text{ kg}^{-1}$. Among the microbial characteristics, microbial biomass C varied from 366 to 478 μ g_c g⁻¹ (mean 410 ± 35) and microbial biomass C/organic C ratio from 2.3 to 3.1% (mean 2.6 ± 0.2 ; Table 2). The biomass C constitutes up to 5% of total organic C^{28} . However, ratios varying from 0.27 to 7.0% have been reported from soils

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Location	pН (1:2.5 H ₂ O)	Clay (%)	CEC $(\mu \text{mol}_e \text{ g}^{-1})$	Organic C $(g \text{ kg}^{-1})$	Total N $(g \; kg^{-1})$
South Andaman					
Collinpur	5.21(0.04)	19(2)	216 (22)	14.6 (1.6)	1.37(0.16)
Shoal Bay	5.32(0.08)	23(2)	231(21)	19.8(1.3)	1.83(0.24)
Crikabad	6.04(0.05)	24(3)	212(19)	14.3(2.2)	1.32(0.17)
Namunagarh	5.71(0.09)	26(2)	232 (21)	16.9(1.9)	1.42(0.11)
Middle Andaman					
Baratang	5.43(0.05)	21(4)	236 (24)	17.1(2.7)	1.62(0.14)
Kadamtala	5.62(0.08)	24(3)	248 (23)	14.6 (3.0)	1.31(0.18)
Betapur	6.05(0.08)	20(3)	222(24)	14.1(1.7)	1.32(0.12)
Nimbutala	5.27(0.07)	24(4)	253(26)	14.2(1.7)	1.38(0.12)
North Andaman					
Kalighat	5.62(0.05)	20(3)	261(23)	15.6(2.1)	1.42(0.21)
Austin Creek	5.36(0.06)	21(5)	268 (24)	16.2(1.4)	1.51(0.22)
R. K. Puram	5.42(0.08)	23(5)	221(16)	16.5(2.1)	1.50(0.16)
Paschim Sagar	5.63(0.08)	25(3)	228 (14)	16.3(2.5)	1.48(0.16)
Little Andaman					
Nethaji Nagar	5.21(0.05)	24(3)	233(21)	16.0(1.6)	1.46(0.13)
Harbinder Bay	5.20(0.08)	26(3)	241(15)	13.9(1.7)	1.28(0.21)
Vivekanandapuram	5.62(0.08)	24(4)	262(23)	15.3(2.1)	1.41(0.21)
Dugong Creek	5.56(0.04)	27(4)	260 (26)	14.6(2.7)	1.43(0.21)
$Mean \pm SD$	5.52 ± 0.26	23.2 ± 2.3	239 ± 17	15.6 ± 1.5	1.44 ± 0.13

Table 1. Relevant physico-chemical properties of soils of various mangrove sites in the Andamans

Values in parentheses indicate standard error of mean.

Table 2. Microbial properties of soils of various mangrove sites in the Andamans

	C_{MIC} ^a	N_{MIC} ^b	$qCO2$ mg $CO2$	C_{MIC} /organic C	
Location	$(\mu g g^{-1})$	$(\mu g \ g^{-1})$	(g biomass C^{-1} d ⁻¹	$(\%)$	$C_{\rm MIC}/N_{\rm MIC}$
South Andaman					
Collinpur	366(49)	32(3.1)	39.6	2.5	11.4
Shoal Bay	462 (37)	39(3.0)	47.2	2.3	11.8
Crikabad	444 (18)	37(2.4)	51.2	3.1	12.0
Namunagarh	450 (18)	34(1.8)	43.8	2.7	13.2
Middle Andaman					
Baratang	478 (37)	33(2.0)	48.5	2.8	14.5
Kadamtala	402(14)	36(2.2)	38.8	2.7	11.2
Betapur	410 (23)	35(1.9)	34.9	2.9	11.7
Nimbutala	371 (14)	32(1.6)	41.0	2.6	11.6
North Andaman					
Kalighat	383 (21)	32(2.4)	38.1	2.4	12.0
Austin Creek	412 (31)	35(3.6)	37.9	2.5	11.8
R. K. Puram	414 (36)	32(3.4)	39.4	2.5	12.9
Paschim Sagar	421 (24)	35(2.7)	38.0	2.6	12.0
Little Andaman					
Nethaji Nagar	414 (39)	34(2.5)	40.3	2.6	12.2
Harbinder Bay	374 (36)	33(3.6)	40.6	2.7	11.3
Vivekanandapuram	384 (21)	34(4.1)	38.3	2.5	11.3
Dugong Creek	370 (22)	32(2.8)	40.2	2.5	11.6
$Mean \pm SD$	410 ± 35	34 ± 2	41.1 ± 4.4	2.6 ± 0.2	12.0 ± 0.9

^a C_{MIC}, Microbial biomass C; ^b N_{MIC}, Microbial biomass N; Values in parentheses indicate standard error of mean.

across different management systems, sampling times and analytical methods²⁹. The microbial biomass N ranged
from 32 to 39 µg g^{-1} (mean 34 ± 2; Table 2). This is lower
than the range (41–54 µg g^{-1}) reported under moist deciduous and semi-evergreen forests of the Andamans¹⁷, but almost identical to the range $(32-36 \mu g g^{-1})$ reported under secondary tropical forest sites of the Philippines¹².

Among the adenylates (Table 3), AMP ranged from 0.21 to 0.29 nmol g^{-1} (mean 0.25 ± 0.03), ADP from 0.41
to 0.48 nmol g^{-1} (mean 0.44 ± 0.03) and ATP from 2.32

Location	ATP $(nmol g^{-1})$	AMP $(nmol g^{-1})$	ADP $(nmol g^{-1})$	AEC	ATP/C _{MIC} (µmol g^{-1})
South Andaman					
Collinpur	2.50(0.14)	0.29(0.04)	0.47(0.08)	0.84	6.8
Shoal Bay	3.02(0.12)	0.24(0.06)	0.46(0.08)	0.87	6.5
Crikabad	3.10(0.32)	0.24(0.05)	0.48(0.06)	0.87	7.0
Namunagarh	3.21(0.21)	0.27(0.05)	0.43(0.04)	0.87	7.1
Middle Andaman					
Baratang	3.22(0.18)	0.26(0.04)	0.48(0.04)	0.87	6.7
Kadamtala	2.52(0.22)	0.28(0.06)	0.48(0.06)	0.84	6.3
Betapur	2.76(0.14)	0.25(0.06)	0.41(0.07)	0.87	6.7
Nimbutala	2.43(0.23)	0.28(0.03)	0.42(0.06)	0.84	6.5
North Andaman					
Kalighat	2.32(0.17)	0.21(0.02)	0.41(0.05)	0.86	6.0
Austin Creek	2.98(0.19)	0.25(0.02)	0.42(0.05)	0.87	7.2
R. K. Puram	2.84(0.21)	0.25(0.03)	0.43(0.03)	0.87	6.8
Paschim Sagar	3.12(0.09)	0.26(0.05)	0.48(0.07)	0.87	7.4
Little Andaman					
Nethaji Nagar	3.11(0.13)	0.21(0.05)	0.45(0.07)	0.88	7.5
Harbinder Bay	2.82(0.15)	0.22(0.04)	0.41(0.04)	0.88	7.5
Vivekanandapuram	3.17(0.19)	0.25(0.05)	0.41(0.05)	0.88	8.2
Dugong Creek	2.81(0.09)	0.21(0.04)	0.43(0.05)	0.87	7.6
Mean \pm SD	2.87 ± 0.29	0.25 ± 0.03	0.44 ± 0.03	0.87 ± 0.01	7.0 ± 0.6

Table 3. Levels of adenylates (ATP, AMP, ADP), AEC and ATP/C_{MIC} ratio of soils of various mangrove sites in the Andamans

Values in parentheses indicate standard error of mean.

to 3.22 nmol g^{-1} (mean 2.87 ± 0.29). Average ATP levels are reported to be 4.2 µmol g^{-1} in grassland soils, 2.1 µmol g^{-1} in forest soils and 1.2 μ mol g^{-1} in arable soils¹⁶. We also observed a positive correlation between biomass C levels and ATP ($r = 0.68$ at $P < 0.001$, $n = 160$) and sum of adenylates ($r = 0.65$ at $P < 0.001$, $n = 160$), which suggested that soils with greater biomass C levels are most likely to possess higher ATP levels. A similar relationship was observed in a large group of soils under different management regimes^{$2,13$}

The mean ATP/microbial biomass C ratio (Table 3) was 7.0 ± 0.6 µmol g^{-1} (range 6.0–8.2; Table 3). This is markedly lower than the average value of 11.7 μ mol g^{-1} and the geometric mean of 10.5 µmol g^{-1} reported in a wide range of soils of the northern and southern hemispheres under diverse management regimes^{2,9}. It is pertinent to note that ATP levels in these studies were determined using the enzymatic luciferin/luciferase system. However, in recent studies wherein ATP was determined using the DMSO extractant, ATP/microbial biomass C ratio ranged between 3.1 and 5.2 µmol g^{-1} in secondary tropical forest sites¹²,
between 3.2 and 8.9 µmol g^{-1} in soils amended with glucose⁴ and between 4.1 and 5.6 µmol g^{-1} in wet tropical forests of the Andamans¹⁷. In the present study also, ATP was determined using the DMSO extractant. The mean ATP/microbial biomass C ratio of 7.0 ± 0.6 µmol g⁻¹ in our mangrove soils is close to the average value of 8.7 μ mol g^{-1} observed in arable soils³⁰ and almost identical to the mean ATP/biomass C ratio of 7.1 umol g^{-1} reported recently in a

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wide range of forests, grasslands and arable soils 13 . They have excluded the possibility of incomplete extraction of the added ATP to be the major reason for such lower ATP/microbial biomass C ratios because of the high extraction efficiency (between 90 and 95%) of the alkaline DMSO extractant. Therefore, the most plausible explanation for the relatively low average ATP/biomass C ratio in the mangrove soils compared to those of Jenkinson⁹ and Contin *et al*,² could be differences in the soil microbial community structure¹³. Probably, the microbial community structure of our mangrove soils is dominated by fungi and microorganisms with large microbial biomass C/N ratio. The microbial biomass C/N ratio is considered to be an indicator of the relative proportion of fungi to bacteria³¹. Consequently, wider ratios (range 11.2–14.5, mean $12.0 \pm$ 0.9; Table 2) indicate that fungi dominated these mangrove soils compared to bacteria. Similar observations were made in forest floor layer and acidic forest A horizon dominated by fungi³² and microorganisms with a large
biomass C/N ratio³³. They^{32,33} attributed this to higher C availability coupled with relatively low N availability due to which the production of enzymes involved in the metabolic pathways producing ATP is inhibited. Earlier reports on distribution of microorganisms in mangrove soils and waters from the Southern region along the coast of the Andaman Sea and the Gulf of Thailand, also indicate that fungi dominate bacteria and algae³⁴. About 60 fungal species were identified, out of which Aspergillus sp., Penicillium sp., Trichoderma sp., Fusarium sp. and

Rhizoctonia sp. were found to be the most frequent in mangrove soils of the Andamans³⁵.

AEC, which indicates the energy status of soil microorganisms ranged from 0.84 to 0.88 (mean 0.87; Table 3). AEC levels ranging from 0.67 to 0.74 under secondary tropical forest sites of the Philippines¹² and those ranging from 0.85 to 0.87 under the moist deciduous and semievergreen forests of the Andamans¹⁷ have been reported. Nevertheless, values ranging from 0.3 to 0.9 have been observed in a wide range of soils⁶. AEC levels >8.0 , similar to those observed in our mangrove soils, have been described in soils where majority of the microorganisms are probably dormant^{10,36}. It is also plausible that highly active microorganisms have large metabolic quotient $(aCO₂)$ coupled with large ATP/biomass C ratios and high AEC^{37} . However, no significant relationship between these three indices has been found in soils 13 . Though significant correlation between AEC and qCO₂ has been observed³⁶, the relationship between these two in our mangrove soils was non-significant. We also did not observe any significant relationship between AEC and individual parameters like pH, CEC, clay, organic C, total N, etc. Therefore, it is still not clear as to how soil microorganisms maintain such high AEC levels, similar to actively growing microorganisms in vitro⁴. Nevertheless, it has been hypothesized that the survival strategy of soil microorganisms is based on a resting population expending energy to maintain a state of metabolic alertness for immediate use of any exogenous substrate³⁸. Therefore, an unknown combination of different factors like quantity and quality of soil organic matter, texture, pH, etc. seems to influence AEC levels in soils 39 .

Overall, the mean ATP levels in our mangroves soils $(0-15 \text{ cm})$ was $2.87 \pm 0.29 \text{ \mu}$ mol g⁻¹ and the mean ATP/ microbial biomass C ratio was 7.0 ± 0.6 µmol g⁻¹, considerably lower than the worldwide average of $10-12 \mu$ mol g⁻¹ ATP g^{-1} biomass C observed in a wide range of soils^{2,9}. This apparent discrepancy is most likely due to a changed microbial community structure possibly dominated by fungi and microorganisms with large microbial biomass C/N ratio. The mean AEC level of 0.87 is within the range of 0.3–0.9 observed in a large group of soils⁶. However, AEC values >8.0 indicate that majority of microorganisms in our mangrove soils are probably dormant. It also needs to be emphasized that the data presented are from soils sampled before the tsunami struck the shores of the Andaman and Nicobar Islands on 26 December 2004. Nevertheless, it can be presumed that variation in the adenylate and AEC levels between the pre- and post-tsunami soil samples would be minimum due mainly to the fact that mangroves are frequently inundated by sea water during periods of high tide. However, at sites that have been permanently submerged post-tsunami, variation in the levels of adenylates, ATP/microbial biomass C ratio and AEC from those reported by us is a distinct possibility.

- 1. Conklin, A. R. and MacGregor, A. N., Soil adenosine triphosphate: extraction, recovery and half-life. Bull. Environ. Contam. Toxicol., 1972, 72, 296-300.
- 2. Contin, M., Todd, A. and Brookes, P. C., The ATP concentration in the soil microbial biomass. Soil Biol. Biochem., 2001, 33, 701-704.
- \mathbf{R} Contin, M., Corcimaru, S., De Nobili, M. and Brookes, P. C., Soil Biol. Biochem., 2000, 32, 1219-1225.
- 4. Joergensen, R. G. and Raubuch, M., Adenylate energy charge of a glucose-treated soil without adding a nitrogen source. Soil Biol. Biochem., 2002, 34, 1317-1324.
- 5. Jenkinson, D. S. and Ladd, J. N., Microbial biomass in soil: measurement and turnover. In Soil Biochemistry (eds Paul, E. A. and Ladd, J. N.), Marcel Dekker, New York, vol. 5, 1981, pp. 415-471.
- 6. Brookes, P. C., Tate, K. R. and Jenkinson, D. S., The adenylate energy charge of the soil microbial biomass. Soil Biol. Biochem., 1983. 15. 9-16.
- 7. Chapman, S. J., Fall, J. and Atkinson, D. E., The adenylate energy charge in Escherichia coli during growth and starvation. J. Bacteriol., 1971, 108, 1072-1086.
- 8. Jenkinson, D. S., Davidson, S. A. and Powlson, D. S., Adenosine triphosphate and microbial biomass in soil. Soil Biol. Biochem. 1979, 11, 521-527.
- 9. Jenkinson, D. S., The determination of microbial biomass carbon and nitrogen in soil. In Advances in Nitrogen Cycling in Agricultural Ecosystem (ed. Wilson, J. T.), CABI, Wallingford, 1988, pp. 368-386.
- 10. Brookes, P. C., Estimation of the adenylate energy charge in soils. In Methods in Applied Soil Microbiology and Biochemistry (eds Alef, K. and Nannipieri, P.), Academic Press, London, 1995, pp. $204 - 213$.
- 11. Joergensen, R. G. and Raubuch, M., Adenylate energy charge and ATP-to-microbial biomass C ratio in soils differing in the intensity of disturbance. Soil Biol. Biochem., 2003, 35, 1161-1164.
- 12. Salamanca, E., Raubuch, M. and Joergensen, R. G., Relationships between soil microbial indices in secondary tropical forest soils. Appl. Soil Ecol., 2002, 21, 211-219.
- 13. Dyckmans, J., Chander, K., Joergensen, R. G., Priess, J., Raubuch, M. and Sehy, U., Adenylates as an estimate of microbial biomass C in different soil groups. Soil Biol. Biochem., 2003, 35, 1485-1491.
- 14. Ross, D. J., Tate, K. R., Cairns, A. and Meyrick, K. F., Fluctuations in microbial biomass indices at different sampling times in soils from tussock grasslands. Soil Biol. Biochem., 1981, 13, 109-114.
- 15. Ross, D. J., Tate, K. R., Cairns, A. and Pansier, E. A., Microbial biomass estimation in soils from tussock grasslands by three biochemical procedures. Soil Biol. Biochem., 1980, 12, 375-383.
- 16. Denobili, M., Diaz-Ravina, M., Brookes, P. C. and Jenkinson, D. S., Adenosine 5'-triphosphate measurements in soils containing recently added glucose. Soil Biol. Biochem., 1996, 28, 1099-1104.
- 17. Dinesh, R., Ghoshal Cahudhuri, S., Ganeshamurthy, A. N. and Dey, C., Changes in soil microbial biomass and their relationships following deforestation and cultivation in wet tropical forests. Appl. Soil Ecol., 2003, 24, 17-26.
- 18. Dagar, J. C. and Sharma, A. K., Multiple viviparity in mangroves. J. Andaman Sci. Assoc., 1989, 5, 72-73.
- 19. Nelson, D. W. and Sommers, L. E., Total carbon, organic carbon and organic matter. In Methods of Soil Analysis, Part 2, Chemical and Microbiological Properties (eds Page, A. L. et al.), Agron. 9, ASA, SSSA, Madison, WI, 1982, 2nd edn, pp. 539-579.
- 20. Bremner, J. M. and Mulvaney, C. S., In Nitrogen-Total. Methods of Soil Analysis, Part 2, Chemical and Microbiological Properties (eds Page, A. L. et al.), Agron. 9, ASA, SSSA, Madison, WI, 1982, 2nd edn, pp. 595-624.
- 21. Gee, G. W. and Bauder, J. W., Particle-size analysis. In Methods of Soil Analysis, Part 1, Physical and Mineralogical Methods (ed. Klute, A.), 1986, Agron. Monograph, vol. 9, 2nd edn, pp. 383-411
- 22. Gillman, G. P., A proposed method for the measurement of exchange properties of highly weathered soils. Aust. J. Soil Res., 1979. 17. 129-139.
- 23. Vance, E. D., Brookes, P. C. and Jenkinson, D. S., An extraction method for measuring soil microbial biomass C. Soil Biol. Biochem., 1987, 19, 703-707.
- 24. Joergensen, R. G., The fumigation-extraction method to estimate soil microbial biomass: extraction with 0.01 M CaCl₂. Agribiol. Res., 1995, 48, 319-324.
- 25. Joergensen, R. G. and Mueller, T., The fumigation-extraction method to estimate soil microbial biomass: calibration of the K_{EN} value, Soil Biol, Biochem., 1996, 28, 33-37.
- 26. Dyckmans, J. and Raubuch, M., A modification of a method to determine adenosine nucleotides in forest organic laver and mineral soils by ion-paired reversed-phase high performance liquid chromatograph. J. Microbiol. Methods, 1997, 30, 13-20.
- 27. Atkinson, D. E. and Walton, G. M., Adenosine triphosphate conservation in metabolic regulation. J. Biol. Chem., 1967, 242, 3239-3241.
- 28. Sparling, G. P., Ratio of microbial biomass carbon to soil organic carbon as a sensitive indicator of changes in soil organic matter. Aust. J. Soil Res., 1992, 30, 195-207.
- 29. Anderson, T.-H. and Domsch, K. H., Ratios of microbial biomass C to total organic C in arable soils. Soil Biol. Biochem., 1989, 21, $471 - 479$
- 30. Martens, R., Estimation of ATP in soil: extraction methods and calculation of extraction efficiency. Soil Biol. Biochem., 2001, 33, 973-982
- 31. Anderson, T.-H. and Domsch, K. H., Quantities of plant nutrients in the microbial biomass of selected soils. Soil Sci., 1980, 130, $211 - 216$
- 32. Blagodatskaya, E. V. and Anderson, T.-H., Interactive effects of pH and substrate quality on the fungal-to-bacterial ratio and qCO₂ of microbial communities in forest soils. Soil Biol. Biochem., 1998 30 1269-1274
- 33. Joergensen, R. G. and Castillo, K., Interrelationships between microbial and soil properties in young volcanic ash soils of Nicaragua. Soil Biol. Biochem., 2001, 33, 1581-1589.
- 34. Chalermpongse, A. and Tappipidh, W., In Proceedings of the Seminar on Microbial Aspects of Nutrient Cycling in Mangrove Environments, UNDP/UNESCO (RAS/79/002), Manila, the Philippines, 1985, p. 47.
- 35. Chauhan, S. K., Tyagi, V. K. and Nagar, M. L., J. Indian Bot. Soc., 1980, 58, 281-285.
- 36. Chander, K. C., Dyckmans, J., Joergensen, R. G., Meyer, B. and Raubuch, M., Different sources of heavy metals and their longterm effects on soil microbial properties. Biol. Fertil. Soils, 2001, 34. 241-247.
- 37. Nannipieri, P., Grego, S. and Ceccanti, B., Ecological significance of the biological activity in soil. In Soil and Biochemistry (eds Bollag, J. M. and Stotzky, G.), Marcel Dekker, New York, 1990, vol. 6, pp. 293-355.
- 38. De Nobili, M., Contin, M., Mondini, C. and Brookes, P. C., Soil microbial biomass is triggered into activity by trace amounts of substrate. Soil Biol. Biochem., 2001, 33, 1163-1170.
- 39. Ciardi, C., Ceccanti, B., Nannipieri, P., Casella, S. and Toffanin, A., Effect of various treatments on contents of adenine nucleotides and RNA of Mediterranean soils. Soil Biol. Biochem., 1993, 25. 739-746.

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Biosorption of metals from contaminated water using seaweed

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Heavy metals are major pollutants in marine, lake and groundwaters as well as in industrial and even treated effluents. Biosorption, an inexpensive and reliable method to remove cadmium and lead ions from solution using dry seaweed biomass as adsorbents, was investigated. Sargassum wightii exhibited maximum metal uptake at pH 4-5 and the value ranged from 18% to 29% of dry biomass. The kinetics of metal adsorption was fast with $70-80\%$ taking place within 30 min. Based on these results, a biobattery involving perforated columns packed with pulverized dry biomass of S. wightii was designed, which could remove metals in the range of 50-97% from a multi-metal ion solution within two and a half hours. The mechanism of metal sorption by seaweeds and the advantages of the present design of seaweed columns are discussed in the light of ecofriendly and cost-effective approach for effluent treatment.

Keywords: Biobattery, biosorption, effluent treatment, heavy metals, Sargassum wightii.

HEAVY metals can be extremely toxic as they damage nerves, liver, kidney and bones, and also block functional groups of vital enzymes¹. Stringent environmental legislation and powers of the authoritative bodies established to enforce these regulations are increasing the demand for new technologies to remove metal from wastewater. For more than a decade, researchers have been looking for cheaper and more effective methods to remediate heavy metal-contaminated waters and reduce the growing publichealth risk. Biosorption is proven to be quite effective at removing metal ions from contaminated solution in a low-cost and environment-friendly manner². The major advantages of biosorption over conventional treatment methods include low cost, high efficiency of metal removal from dilute solution, minimization of chemical and/or biological sludge, no additional nutrient requirement, regeneration of biosorbent and the possibility of metal recovery³.

Bacteria⁴, fungi⁵, marine algae^{6,7}, etc. have been studied for their heavy metal uptake capacities and suitability to be used as development of biosorbents. Biosorptive capacities of seaweeds, activated carbon and natural zeolites have been evaluated and are comparable to those of synthetic ion-exchange resins⁶. Marine macro-algae are harvested or cultivated in many parts of the world and are therefore readily available in large quantities for the development of highly effective biosorbent materials. This study inves-

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